The effect of tert-butyl hydroperoxide on hepatic transcriptome expression patterns in the striped sea bream (Lithognathus mormyrus; *Teleostei***)**

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Abstract

The study was aimed at examining the effects of tert-butyl hydroperoxide (tBHP) on hepatic transcriptome expression patterns of the teleost fish *Lithognathus mormyrus*. tBHP is an organic hydro-peroxide, widely used as a model pro-oxidant. It generates the reactive oxygen species (ROS) tert-butoxyl and tert-butylperoxyl. Complementary DNAs of tBHP-treated vs control fish were applied onto a previously produced cDNA microarray of ~ 1500 unique sequences. The effects of the tBHP application were demonstrated by leukocyte infiltration into the liver and by differential expression of various genes, some already known to be involved in ROS-related responses. Indicator genes of putative ROS effects were: aldehyde dehydrogenase 3A2, Heme oxygenase and the hemopexin-like protein. Putative indicators of transendothelial leukocyte migration and function were: p22phox, Rac1 and CD63-like genes. Interestingly, 7-dehydrocholesterol reductase was significantly down-regulated in response to all treatments. Several non-annotated genes revealed uniform directions of differential expression in response to all treatments.

 Keywords: *Reactive oxygen species , aqueous medium , liver , fi sh , tert-butyl hydroperoxide , gene expression*

Introduction

Concerted expression of genes underlies almost all biological processes. Consequently, transcriptome expression patterns are widely used as increasingly important research and diagnostic tools across biology and bio-medicine [1,2]. The present study was aimed at elucidating the effect of the model reactive oxygen species (ROS) precursor, tert-Butyl Hydro Peroxide $(tBHP)$ (CH₂)3COOH, on hepatic transcriptome expression patterns in the teleost fish *Lithognathus mormyrus*.

ROS are chemically active species of oxygen, harmful to a variety of biological macromolecules and processes in living organisms. They are products of incomplete reduction which occurs during normal reduction of molecular oxygen. They are produced also through metabolism of xenobiotic ROS precursors. ROS

interfere with cellular functions through impairing nucleic acids, proteins and lipids. Oxidative stress in response to exposure to ROS occurs when the adaptive detoxification capacity of the cellular anti-oxidizing mechanisms is insufficient $[3,4]$. Defense mechanisms against oxidative stress include up-regulation of anti-oxidation-related genes, neutralization of ROS by small anti-oxidant scavenger molecules [5] and elimination of damaged proteins through the ubiquitin-proteasome [3] and autophagy [6] pathways. ROS were shown also to be signalling molecules participating in the maintenance of their own homeostasis and in other regulatory pathways $[7-11]$.

Several pollutant types can alter the equilibrium between pro- and anti-oxidants leading to oxidative stress [4,12]. Oxidative stress stimulated by pollutants has been shown to occur in fish $[13]$. Liver, gills and

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intestine are interesting organs in the environmental context of oxidative stress. The liver is a centre of xenobiotic metabolism and the other organs are located at the interface between the environment and the fish body. This study was concentrated on the liver.

tBHP is an organic hydro-peroxide widely used experimentally as a model pro-oxidant. It has been applied to mammals and mammalian cells $[14-17]$ and also to fish, as a dissolved compound in the ambient water and as an injected compound [18-20]. tBHP metabolism in the liver includes several iron ion related reactions, involving both free ions and ions acting through the cytochrome P450 heme. Two free radical species are produced, tert-butylperoxyl and tert-butoxyl. The latter can initiate lipid peroxidation [21].

It is hypothesized that increased levels of ROS are produced as a result of tBHP exposure and may cause changes in the expression of specific genes. These changes are hypothesized to be related to defense mechanisms against free radicals or to toxic effects caused by the free radicals.

The utilization of the hepatic transcriptome expression profile of the fish *L. mormyrus* as an environmental biomarker is continuously studied in our laboratory [22] and an hepatic cDNA microarray of *L. mormyrus*, cloned from a multi-pollutant affected transcriptome, has served as an expression evaluation tool [22]. tBHP has not yet been identified in natural environments, in spite of its industrial utilization [23]. Therefore, it does not directly mimic environmental situations. However, being a prooxidant, it can produce at least part of the ROSrelated cell response repertoire. Its experimental application was therefore expected to contribute to the broad perspective of ROS toxic effects on hepatic expression patterns in *L. Mormyrus*, hence, to be relevant to the potential utilization of this fish as an environmental sentinel.

Materials and methods

Fish sampling and processing

L. mormyrus individuals were sampled alive by gillnetting in a clean habitat along the Mediterranean coast of Israel. Winter sampling was aimed at avoiding catching fish during their reproductive season. Two exposure experiments to tBHP were conducted. Fish were kept in the laboratory maintenance facilities 1 or 2 months prior to experiments 1 and 2, respectively. Flow through seawater at $20-21^{\circ}$ C was supplied. Length and weight of the fish were measured at the termination of each experiment, followed by immediate sacrifice by decapitation and dissection of the liver and the gonads. Livers were divided into smaller aliquots, snap frozen in liquid nitrogen and stored at -80° C. Gonads were fixed in 4% formaldehyde. The fish reproductive activity was determined by microscopic observation of fixed gonads smeared between two slides, measuring the average ova diameter of 20 oocytes. Being a proteandrous hemaphrodite, *L. mormyrus* ovotestes always contained at least primordial oogonia [24]. Hence, individuals containing solely primordial oogonia were selected for further analysis.

Total hepatic RNAs were extracted from livers a few days after freezing using the EZ II kit (Biological Industries, Beit Haemek, Israel) according to the manufacturers' instructions. Messenger RNA was isolated from the total RNA preparations using the PolyATtract kit (Promega, Madison, WI, USA). The quality of produced RNAs was evaluated by electrophoresis on a 1% agarose-formaldehyde gel and their concentration was evaluated by spectrophotometry (Nanodrop, Nanodrop Technologies, Wilmington, DE).

Two experiments were conducted:

Experiment 1. Three groups with five fish in each of them were injected with three tBHP doses, 0.5, 5 and 50 mg/kg body weight/injection. Three injections were applied to each fish with 3-day intervals between consecutive injections. A control group of five fish was sham-injected with saline, the tBHP carrier. Both the control and the treated fish were sacrificed 3 days after the last injection. Hence, fish were exposed to tBHP during 9 days. The injected concentrations were aimed at applying a broad range of tBHP sub-lethal doses, initially determined in previous tBHP exposure studies cited in the introduction.

Experiment 2. Two groups with 18 fish in each of them were injected with two tBHP doses of 5 and 30 mg/kg body weight/injection. Five consecutive injections were applied daily to each fish. The first injection of the high dose contained 50 mg/kg body weight, but this was reduced to 30 mg/kg body weight due to mortality of two fish several hours after the first injection. A control group of 18 fish was sham-injected by similar volumes of saline. Six fish of each of the treatment and control groups were sacrificed at three time points: 2, 20 and 68 h after the last injection. Hence, fish were exposed to tBHP during $4-7$ days.

Fish characteristics of both experiments are presented in Table I. They reveal no statistically significant differences in fish size among samples and an absence of reproductive activity, judged by their oocyte diameter, according to Funkenstein et al. [25].

Microarray hybridization and analytical procedures

The construction procedure of the utilized microarray followed Auslander et al. [22]. Briefly, cDNAs were

*Average ± standard deviation.

cloned from hepatic mRNA of multi-pollutant exposed fish. tBHP was not one of the inducing compounds. The cloned cDNAs were amplified by PCR and the cleaned PCR products were printed on a treated glass slide. In addition, the cDNAs were sequenced and annotated using the BLASTN and BLASTX programmes [26]. Additional functional information was provided by the Gene Ontology (GO) database [27] and the Kyoto Encyclopedia of Genes and Genomes (KEGG) database [28] using the Blast2Go software [29] and the KEGG bio-informatics tools (http:// www.genome.jp/kegg/). Functional characterization of each clone was summarized based on the combined information. DNA fragments which share sequences were assembled and the resulting sequence was designated here *unique clone*, representing the sequence of a specific transcript. The microarray platform characteristics were deposited in the Gene Expression Omnibus (GEO) database of the American National Center for Biotechnology Information (NCBI) (GEO platform accession number - GPL 5351).

The hybridization procedure is composed of the labelling protocol of the target RNAs extracted from the fish livers, their dual hybridization on the microarray slides and the imaging of the resulting labelled slides. The three procedures are described in details in Yudkovski et al. [30]. Briefly, labelling of hepatic mRNA populations with the fluorescent dyes Cy3 and Cy5 was performed using the amino-allyl indirect cDNA labelling method. Each target mRNA population was reverse transcribed in the presence of anchored oligo dT primer and dNTP mix in which the dTTP had been replace by an amino allyl-dUTP/dTTP mixture (4:1). The resulting cDNAs were chemically coupled to an aliquot of the Cy3- or Cy5-mono reactive N- hydroxyl succinimidyl ester-derivatized fluorescent dye via the reactive amino allyl groups. Labelling efficiency was evaluated by spectrophotometry at 260 nm for cDNA estimation and 550 nm (Cy3) and 650 nm (Cy5) for determining labelling intensity.

Microarray slides were incubated with a prehybridization solution [30] to block non-specific DNA interactions, thus reducing background fluorescence followed by washing. Dual-labelled solutions for binary comparisons between the two experimental conditions were prepared by mixing equal amounts of the two labelled cDNA populations to be compared, followed by a 1:1 dilution with a hybridization solution [30]. The heat denatured hybridization mixture was layered onto the microarray, covered and incubated overnight at 42° C in a humidified hybridization chamber. Following incubation, the slides were washed and dried.

GenePix Pro 6.1 software (MDS Analytical Technologies, Sunnyvale, CA) was used for slide imaging. Hepatic significant differentially expressed genes were determined by the LIMMA software [31,32] assisted by the LIMMA user guide. Two spotrelated parameters were calculated: $M = log_2$ (Cy3/ Cy5); the $log₂$ expression ratios for each dual labelled spot on each array, and $A = (log_2 (Cy5[*]Cy3))/2$; the average fluorescence intensity across the two dyes. Cy3 and Cy5 are the two normalized values of the fluorescence signals. LIMMA provided also the average A, calculated across each experiment. Only spots with $A > 8.5$ (362 arbitrary fluorescence units out of a maximum of 65 536) and which agreed with the GenePix Pro 6.1 eligibility criteria were included in the analysis. LIMMA analytical procedure was modified to conform to the background characteristics and the distribution of the differentially expressed genes. Hence, the Cy5 and Cy3 intensities within each slide were normalized using the print-tip LOW-ESS method. No background subtraction was applied by LIMMA, as the area under each spot was protected from a non-specific fluorescent signal, revealing lower signal than the background when not specifically hybridized. LIMMA was also used to statistically test the hepatic differentially expressed genes in each defined sample ($M \neq 0$; $p < 0.05$). *P*-values were adjusted for multiple testing corrections [33]. The differentially expressed clones were further selected by applying $M \ge 0.5$ threshold. This threshold was aimed at minimizing false-positive M values resulting from suspected weak differential expression, even if statistically valid, leaving only the more stringent differential expressions.

Experimental design of the hybridizations

RNA preparations resulting from the two exposure experiments were labelled and hybridized on the cDNA microarray. The hybridization designs of Experiments 1 and 2 are described in Figure 1. The design is composed of biological replicates. Each replicate is actually a complete experiment, including randomly selected individual fish from each of the treatments and controls. Hence, the number of the biological replicates in each experiment equals the number of fish in each experimental group, five replicates in Experiment 1 and six in Experiment 2. The design of Experiment 1 (Figure 1A) included intensive dye-swap which contain four technical replicates for each hybridized cDNA pair. The total number of slides in Experiment 1 comprised: 4 technical replicates of each slide \times 3 hybridized fish pairs in each biological replicate \times 5 biological replicates = 60 arrays. When examining the added information of

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Figure 1.Hybridization designs of both experiments. Circles represent randomly selected fish from each of the treatments, designated by the tBHP concentrations in mg/kg body weight. Ellipses represent randomly selected sham-injected fish. Each arrow represents dual-labelled slide of cDNA preparations of a fish pair. Arrowhead represents Cy3 and arrow tail, Cy5. Each figure (A, B1 and B2) represents a biological replicate. (A) Experiment 1, included four dye-swapped technical replicates per hybridized labelled cDNAs of a fish pair. The total number of slides in Experiment 1 was: 4 technical replicates \times 3 hybridized fish pairs in a biological replicate \times 5 biological replicates = 60 arrays. (B1 and B2) Experiment 2, The total number of slides in Experiment 2 was: 2 hybridized fish pairs in a biological replicate \times 6 biological $replicates = 12$ arrays. Alternate labelling of half of the biological replicates (B1 or B2) was used, partially compensating the dye effect.

each technical replicate of Experiment 1, it occurred to us that they did not provide additional information justifying the effort. Therefore, Experiment 2 was differently designed. No technical replicates were applied and potential dye effects were partially compensated by the alternate labelling of half of each biological replicate (Figures 1B1 and 1B2). The total number of slides in Experiment 2 was: 2 hybridized pairs in a biological replicate \times 6 biological replicates $= 12$ arrays.

Relative qRT-PCR

Relative expression levels of target transcripts (Table II) between selected RNA populations of individual treated fish and their mutually hybridized individual controls were evaluated using reverse transcription coupled to relative real-time PCR, according to Pfaffl [34]. The used primer pairs are presented in Table II. Total RNAs of individual fish were serially diluted to 1, 0.5, 0.25 and 0.125 μg. Each of them was reverse transcribed at 42°C for 1 h using MMLV reverse transcriptase (Promega, Madison, WI, USA) and reverse primers, according to Sambrook and Russell [35]. PCR reactions were performed using the four reverse transcriptase solutions of the serially diluted RNAs as templates. Each reaction was performed in triplicate with the GeneAmp 5700 PCR thermocycler (Applied Biosystems, Foster City, CA) (one cycle at 50° C, 2 min; one cycle at 95° C, 10 min; 40 cycles at 95° C, 15 s and 60° C, 1 min). Each PCR reaction was

conducted with a volume of 25 μl composed of 12.5 μl SYBR Green mix (Applied Biosystems), 0.2 μM of each of the PCR primers and a 2 μl aliquot sample from the reverse transcription mixture. A linear regression equation relating the crossing-point (CP) of the PCR cycle to the log of the four dilutions was calculated for each PCR-reacted RNA population. CP is the PCR cycle in which the fluorescent intensity crosses the threshold of detection. The equation slope was used to ensure a reasonably similar PCR efficiency of the two RNA populations compared. ΔCP is the difference between the CPs of two compared populations at identical dilution. Their average across the four dilutions was used to express the difference between the two compared beta-actin normalized RNA populations. Both M-values and Δ CPs represent log₂ expression ratio between compared RNA extracts. Although evaluated by different methods, they can be qualitatively compared, demonstrating expression trends. The Δ CPs were compared to the corresponding M values taken from the microarray results (Figure 2).

Histology

Liver tissue pieces from all fish of Experiment 2 were fixed in Bouin's reagent, embedded and stained by haematoxylin and Eosin according to Stevens [36]. Pathological structures were identified according to Feist et al. [37].

Results

Analysis of hybridized slides

Experiment 1. All the three applied tBHP doses were sub-lethal. One hundred and thirteen annotated unique clones (Table III) and 56 non-annotated ones (not shown) fulfilled our criteria for differential expression with $M \ge 0.5$ in at least one of the treatment groups. M values of redundant clones sharing one unique sequence were averaged. Up-regulation of 28, 45 and 58 clones was observed in the 0.5, 5 and 50 mg tBHP/kg body weight/injection dosed groups, respectively, revealing dose dependency. No such dependency was observed for the down-regulated clones (34, 39, 24 clones in the 0.5, 5 and 50 mg tBHP/kg bwt/injection dosed groups, respectively). Nine days elapsed from the first injection to fish sacrifice. The results of Experiment 1 were deposited in the Gene expression omnibus (GEO) database maintained by the American National Center for Biotechnology Information (NCBI) in dataset accession number GSE19216, sample accession numbers GSM476020, GSM476029 and GSM476030.

Experiment 2. This was aimed at examining the gene expression patterns during shorter time periods than

Accession number		Primer pairs	Microarray results (Experiment 2)	qRT-PCR results	Annotation
DO849783	r	CAGAAGCACTCTGAGCAAATCC GCACACACTGATGCATTCAGC	Up-regulation	4.982	Haemoglobin beta-A chain
DO850833	r	AGACGTTCACCTTCTTACGCC AGAGACTGACCCCTGTTCTCC	Up-regulation	2.245	Betaine-homocysteine methyltransferase
DO849829		GTGTTACTCCTGCCCTGACG TCTCACCACTTTTGAGTTTGAGG	Down-regulation	-1.458	Null
DO849652	r	TGACTTTTGTGTCCCTGACTGG CTGCGAAAAACTGACAAAATCC	Down-regulation	-1.957	Complement component C7
DO850757		CTGTGCAAAATAGATCATTGGC TGCCACTACAACTCATTCTACCTC	Down-regulation	-0.682	7-dehydrocholesterol reductase
DO850885	r	TCAAGCAGTTCCACGACTCC CGACAAAATGCTTTCAGTAGAAGG	Down-regulation	3.247	Ribosomal protein L18a (fragment)

Table II. qRT-PCR of selected genes, carried out for microarray validation.

f and r designate forward and reverse primers, respectively.

those of Experiment 1. Only $4-7$ days elapsed from the first injection to fish sacrifice. The 5 and 30 mg tBHP/kg bwt doses were applied in this experiment and only fish dosed with the higher amount of tBHP and sacrificed after 7 days from first injection elucidated differentially expressed clones (31 clones). All these clones were differentially expressed also in Experiment 1 and the annotated ones are included in Table III. The results of Experiment 2 were deposited in the Gene expression omnibus (GEO) database in dataset accession number GSE19216, sample accession numbers GSM476074 and GSM476112.

Prominent differential expression was revealed in Experiment 1, even after exposure to relatively low

Figure 2. Histological appearance of leukocyte infiltration.

doses for 9 days. In contrast, milder changes of expression were detected in Experiment 2 and only at the highest tBHP dosage group after 7 days of exposure.

The differentially expressed genes were classified to functional groups (Table III). Certain functional groups were characterized by a uniform trend of differential expression. Hence, all the proteolytic enzymes, their inhibitors, the genes related to amino acid metabolism, genes related to energy provision and the signal transduction-related genes were upregulated. The genes related to lipid and cholesterol metabolism were highly represented in the differentially expressed genes and most of them were upregulated. Clones of the complement component of the innate immune system revealed both up- and downregulation. C1q, C3 and C7 were down-regulated and Bf-1 and C9 were up-regulated. Most oxidationreduction enzymes were up-regulated as well as proteins related to heme degradation. Other differentially expressed functional gene groups affected by exposure to tBHP were genes related to sugar metabolism and metal ion homeostasis. The majority of the non-annotated differentially expressed genes revealed constant expression direction across treatments.

Several of the differentially expressed genes can be assigned to ROS metabolism: aldehyde dehydrogenase 3A2, Cytochrome p4501A, warm temperature acclimation-related 65 kDa protein (Wap65)-Hemopexin, Heme oxygenase and Heme oxygenase 1, all of which showed up-regulation. Metallothionein genes revealed down-regulation. NADPH oxidase flavocytochrome b small sub-unit p22phox and Ras-related C3 botulinum toxin substrate 1 (Rac1) are known to participate in the regulation of endothelial intercellular adhesion, part of the process of leukocyte transendothelial migration and CD63-like gene participate in the leukocyte function at the inflammation site. For references to the above functions see the discussion section below.

Several other genes showed tBHP-related differential expression. These were the up-regulated betaine

Table III. Short annotations and M values of genes differentially expressed after exposure to tBHP.

(Continued)

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Table III. *(Continued)*

Exp. 1 and 2 are the two performed experiments. The numbers 0.5, 5, 50 and 30 are injected tBHP doses in mg/kg body weight/injection. Only Experiment 2 data for the time point of 68 h after last injection is presented (see Materials and methods for details of the experiments). Positive M values designate up-regulation; negative M values designate down-regulation and blank cells designate no significant differential expression. Bold M values designate $M \geq |0.5|$.

homocysteine methyltransferase (BHMT), Haemoglobin beta-a chain and ceruloplasmin and the downregulated Ribosomal protein L18a, Hepcidin-like/ Hepcidin precursor and insulin-like growth factor 1.

7-Dehydrocholesterol reductase (7-DHCR) (Accession number DQ850757) and two non-annotated genes (Accession numbers DQ849868 and DQ849829) were prominently attenuated in all the treatment groups in both experiments $(M < -0.9)$.

Quantitative RT-PCR validation of microarray results

Relative qRT-PCR was performed with six selected genes using selected mRNAs from two pairs of mutually hybridized fish from Experiment 2, sacrificed 68 h after last injection. Each pair included a fish dosed by 30 mg/kg body weight and a sham-injected fish. The ΔCP between treated and sham-injected fish of each examined pair were compared to the microarray M values obtained for those fish pairs. The results are presented in Table II. Five out of six qRT-PCRs agreed with the microarray M values.

Liver histology

Several apparent pathological structures were observed in the hepatic histological sections taken from fish of Experiment 2. They included paracristal inclusions, vacuolization, spongiosis, foci of cellular alteration, necrosis, high level of macrophage aggregates, bile

Discussion

tBHP, a model pro-oxidant, was applied through peritoneal injection, a direct effective application route in comparison to exposure through the gills or the intestine. However, it took at least 7 days (high dose and longest exposure duration) in Experiment 2 to detect differentially expressed genes. The 7 days gap between application and its effect is probably composed of the transportation period from the peritoneum to the liver, the period required for the postulated production of ROS from tBHP and the period required for visualization of ROS effects on gene expression. The levels of the assumedly produced ROS was not directly measured.

It was mainly the more prominent differential expression results of Experiment 1 that were used to identify differentially expressed genes. Differential expression of 169 unique clones was observed in this experiment in response to all applied tBHP doses in comparison to the appropriate control, but only the numbers of up-regulated genes corresponded to the tBHP injection dose. It is postulated that this dose responsive up-regulation is related to specific regulative and defensive pathways activated directly or indirectly by elevated ROS levels. Down-regulation may be a result of specific regulative and defensive pathways as well. However, the lack of similar dose response in down-regulated genes indicates a mixture of putative specific ROS-related regulation pathways and non-specific toxicity, both affected by tBHP.

Fish liver contains several types of cells which facilitate regular liver and pancreatic functions. Other cell types such as leukocytes migrate to the liver in pathological situations. The pathological effect of the tBHP injections on the liver was demonstrated in

Table IV. Fish from *Experiment 2* revealing hepatic leukocyte infiltration.

	Sacrifice time (h) tBHP dose (mg/kg bwt/injection) $\#$ of fish [†]	
	0 (Sham injected)	
	5	
	50	3
20	0 (Sham injected)	
	5	2
	50	3
68	0 (Sham injected)	
	5	
	50	2

[†]Out of six fish/group.

Experiment 2 by leukocyte infiltration in all dose- and time-related treatment groups starting 4 days after treatment. Leukocyte infiltration is known as a general response to inflammation, but also to the presence of ROS [38,39].

An attempt was made to find common denominators to certain groups of differentially-expressed genes in relation to their functions and in view of their expression upon tBHP application. Assignment of genes to a functional group does not necessitate functioning in the same pathway or activation by the same regulatory route. Nevertheless, several functional groups of genes elucidated a uniform trend of differential expression, which can be interpreted in the context of ROS effect. Hence, all the elucidated proteolytic enzymes, proteolysis inhibitors, amino acid metabolism-related enzymes and genes which are part of the energy providing system were solely up-regulated. This trend may indicate an activity aimed at preserving homeostasis of properly functioning protein molecules in the liver, while eliminating damaged ones.

Unsaturated lipids are sensitive to oxidative stress and lipid peroxidation causes damage to membranes, low density lipoproteins and free fatty acids. The lipid metabolism-related genes and specifically those which are related to cholesterol metabolism were a prominent component of the differentially-expressed genes. There was no uniform direction for the differential expression, but the majority of genes in this group were up-regulated. Again, broadly speaking, lipid homeostasis may be preserved by the differential expression of these genes, both synthetic and catabolic ones. Interestingly, the only prominently downregulated lipid metabolism-related transcript in response to all tBHP exposures was 7-dehydrocholesterol reductase (7-DHCR). 7-DHCR catalyses the conversion of 7-dehydrocholesterol to cholesterol. We could not find a reasonable explanation for its down-regulation in response to tBHP exposure.

The differentially expressed genes which belong to the innate immune system elucidated a mixed trend of expression. It is possible that changes in the expression level of the innate immune system genes is related to the general inflammation response indicated by the leukocyte infiltration (Figure 2).

It was not possible to distinguish the exact cellular origin of differentially expressed genes using the present experimental procedure. However, the process of leukocyte transendothelial migration was intensively studied in mammals [40,41] and may assist in indicating differentially-expressed genes putatively participating in this process. Basigin (CD147) enhances leukocyte cellular adhesion to endothelial cells by its binding to integrins [42]. It was up-regulated here and may participate in the adhesion stage of leukocyte transendothelial migration. NADPH oxidase flavocytochrome b small sub-unit p22phox is a regulatory sub-unit of NADPH oxidase (Nox) enzymatic complex, a major ROS producer [43]. The ROS produced by Nox are used as part of anti-bacterial defense mechanisms [44]. However, Nox is also part of the complex regulatory pathway of leukocyte transendothelial migration. The Nox system, including p22phox, is induced to produce ROS by genes involved in leukocyte-endothel adhesion, which in turn signals dysfunction of endothelial cell-to-cell adhesion. The later decreased adhesion enables leukocyte transendothelial migration [11].Alternatively, as shown by Djordjevic et al. [45] in human endothelial cells, p22phox levels were elevated by a direct redox-sensitive mechanism. Ras-related C3 botulinum toxin substrate 1 (Rac1) is a small GTPase G signal protein belonging to the Rac sub-family of the Rho GTPases. It functions as a regulator of endothelial inter-cellular adhesion and of cyto-skeletal migration on the actin network, through its participation in the Nox pathway $[11,46]$ Hence, it is involved in leukocyte migration. Its down-regulation here is, on one hand, incompatible with the p22phox up-regulation [11]. However, on the other hand its role in controlling endothelial permeability is still controversial [47].

The up-regulated CD63-like gene may be involved in another leukocyte-related function of protein catabolic processes at the hepatic inflammation sites, as it is involved in targeting mechanisms of pro-neutrophil elastase to its site of action [48].

Certain differentially expressed transcripts are known to be related to ROS metabolism and their patterns of expression further indicate a tBHP effect. Aaldehyde dehydrogenase 3A2, an NAD-dependent oxidoreductase, detoxifies aldehyde toxicity caused in the course of alcohol metabolism or lipid peroxidation. The enzyme was previously used as a biomarker of pre-neoplastic hepatocyte foci in fish $[46, 49]$.

Cytochrome p450 1A (CYP1A) is a monooxygenase, reducing molecular oxygen to water while transferring the other oxygen atom into a variety of organic compounds. CYP1A is induced by several groups of organic pollutants (e.g. dioxins, PCBs and PAHs) and its expression level is one of the most established environmental biomarkers in fish [50–52]. It is known to produce ROS in the course of its normal function. However, oxidative stress led to its inactivation [53–55]. This inactivation was interpreted as a defense mechanism against oxidative stress by the reduction of ROS producing enzymes. Increasing transcript levels of CYP1A in response to tBHP exposure may be explained by potential tBHP inducing activity of CYP1A that is not related to the inactivation of the enzyme or by a yet unknown compensating mechanism, increasing transcript level in response to reduced enzymatic activity.

The heme catabolytic enzymes heme oxygenase, alpha-1-microglobulin bikunin precursor and Wap65- Hemopexin were mostly up-regulated. Graça-Souza et al. [56] and Arruda et al. [57] showed that heme is a pro-inflammatory molecule when present in a suitable concentration, leading to Nox-dependent formation of ROS. Therefore, increasing heme degradation is a probable response to inflammation and the consequent increased ROS level. The fish warm temperature acclimation-related 65 kDa protein (Wap65) is a hemopexin-like gene, increasingly expressed in fish with the elevation of the ambient temperature [58]. Its resemblance to hemopexin indicates an additional function, trapping free heme and its transportation to the liver for its catabolism and iron recycling [59]. Gracey et al. [60] showed increased expression of hemopexin-like protein during fish hypoxia. Aliza et al. [59] suggested this phenomenon as an explanation for the increased hemopexin obtained upon exposure of the fish *Xiphophorus helleri* to copper. Increased level of hemopexin transcript is correlated to increased binding of its protein derivative to heme [61,62]. Binding of the heme–hemopexin complex to hemopexin receptor located on the cell membrane can initiate a sequence of cellular events [63,64]. Heme oxygenase (HO) is the first enzyme in the heme catabolic pathway, producing the catabolytes biliverdin, iron and carbon monoxide. HO is induced during oxidative stress [66] and it is assumed that its products biliverdin and bilirubin are potential antioxidants $[66-68]$.

Metallothionein is a cysteine-rich protein involved in the homeostasis of essential metals, detoxification of toxic metals and free radical scavenging [69]. The transcript of this gene is dramatically induced in response to increased cadmium and zinc and also in response to oxidative and inflammatory factors [70,71]. An oxidative stress response and cadmium response sequence was located in the promoter of the mammalian MT-1 [72,73]. A specific antioxidant response element (ARE) was identified along this sequence [74]. However, an un-explained reduction of metallothionein expression was revealed in Experiment 1 here and by Williams et al. [18] upon fish exposure to tBHP.

Other differentially expressed transcripts which were elucidated only in some of the tBHP treatments included the up-regulated betaine homocysteine methyltransferase (BHMT) and haemoglobin beta-a chain and the down-regulated complement component C7, ribosomal protein L18a, hepcidin-like/hepcidin precursor and insulin-like growth factor 1.

Many of the differentially-expressed genes were nonannotated. However, the vast majority elucidated a uniform expression pattern across treatments, up- or down-regulation, indicating cause and effect relationships between the tBHP exposure and the expression pattern. These genes should be further studied.

The only other fish exposure study to injected tBHP was done by Williams et al. [18]. In that study, a single 5 mg tBHP/kg body weight intraperitoneal

injection was applied to the European flounder, followed by evaluation of gene expression patterns 1, 2, 4, 8 and 16 days after injection. In addition to the different gene composition of their microarrays, different tBHP doses, injection schedules and sacrifice timing make the present results difficult to compare to Williams et al. [18]. Nevertheless, few genes, including haemoglobin beta-a chain, vitellogenin and Cytochrome P4501A, were up-regulated and Metallothionein was down-regulated throughout all treatment groups in both studies. Ribosomal protein L18a, NADPH oxidase flavocytochrome b small sub-unit p22phox and most of the hepcidin clones showed contradicting trends between the two studies.

Possible reasons for inaccuracies and false results in microarray assays have been thoroughly discussed in the literature. They are assigned mainly to crosshybridizations of similar sequences belonging to different genes [75–77]. Real-time PCR, an accurate, sensitive and quantifiable method, is usually used to confirm microarray results. Despite its advantages it may also be biased due to accidental utilization of primers which are suitable to more than one transcript. Only one out of the six transcripts examined here by qRT-PCR showed trends which were not compatible with the microarray results.

The interest in hepatic expression patterns of *L. mormyrus* emerged from its potential utilization as an environmental sentinel. Albeit, tBPH, its application route and the concentrations used did not represent likely interactions with known environmental pollutants. Nevertheless, known environmental pollutants can affect cellular redox status [13] and tBHP injections could therefore mimic some of these redox responses.

In summary, this microarray study of liver cell responses to the pro-oxidant tBPH detected some previously observed ROS-related transcription patterns. It also detected novel differential expression of annotated and non-annotated genes that were not previously known to be associated with ROS challenge. These genes are promising candidates for further study of ROS-related metabolism in fish.

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